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Photolysis of bis(1*H*-tetrazol-5-yl)amine (H₂BTA), a high nitrogen content tetrazole-based energetic compound in water

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1 **ABSTRACT**

2 Tetrazoles have wide industrial applications, notably in the pharmaceutical industry.
3 Tetrazole derivatives, such as bis(1*H*-tetrazol-5-yl)amine (H₂BTA) with a nitrogen content
4 of 82.34%, have recently been considered by the defense industry as high nitrogen
5 composite propellants. Little information is available on their environmental fate.
6 Photodegradation studies under solar simulating conditions showed that H₂BTA was
7 partially degraded in water, while it was completely degraded under UV light at $\lambda = 254$
8 nm. When H₂BTA (0.35 mM) was irradiated with simulated sunlight at pH 3.65, there was
9 a 1-day lag phase before the chemical started to degrade, reaching 43.5% degradation after
10 7 days. However, when pH increased to 5.76, H₂BTA degraded without lag phase,
11 suggesting that an HBTA⁻ anion was involved in the initial degradation of the chemical. 5-
12 Aminotetrazole (5-AT) was identified as a final degradation product and *N*-(1*H*-tetrazol-
13 5-yl)formamide (T(5yl)FA) and 1*H*-tetrazol-5-ylcarbamic acid (T(5yl)CA) as intermediate
14 products. At $\lambda = 254$ nm, H₂BTA (0.35 mM, pH 3.65) disappeared rapidly, resulting in
15 the loss of 94% after 65 min. 5-AT was detected together with several transient products
16 including *N*-(1*H*-tetrazol-5-yl)carbamohydrazonic acid (T(5yl)CHA) and T(5yl)FA.
17 Kinetic studies and products analysis revealed that H₂BTA photodegraded via two initial
18 routes. One route (*a*) marked by the initial loss of HN₃ and another (*b*) marked by the initial
19 loss of N₂. Route *a*) was characteristics for irradiation with simulated sunlight; however,
20 routes *a*) and *b*) proceeded simultaneously under UV light. 5-AT eventually degraded to
21 presumably give N₂ and/or HN₃ using UV light. Understanding the photodegradation
22 pathway of H₂BTA under simulated sunlight can help in providing the basis for natural
23 attenuation assessment of the chemical in contaminated aquatic environments.

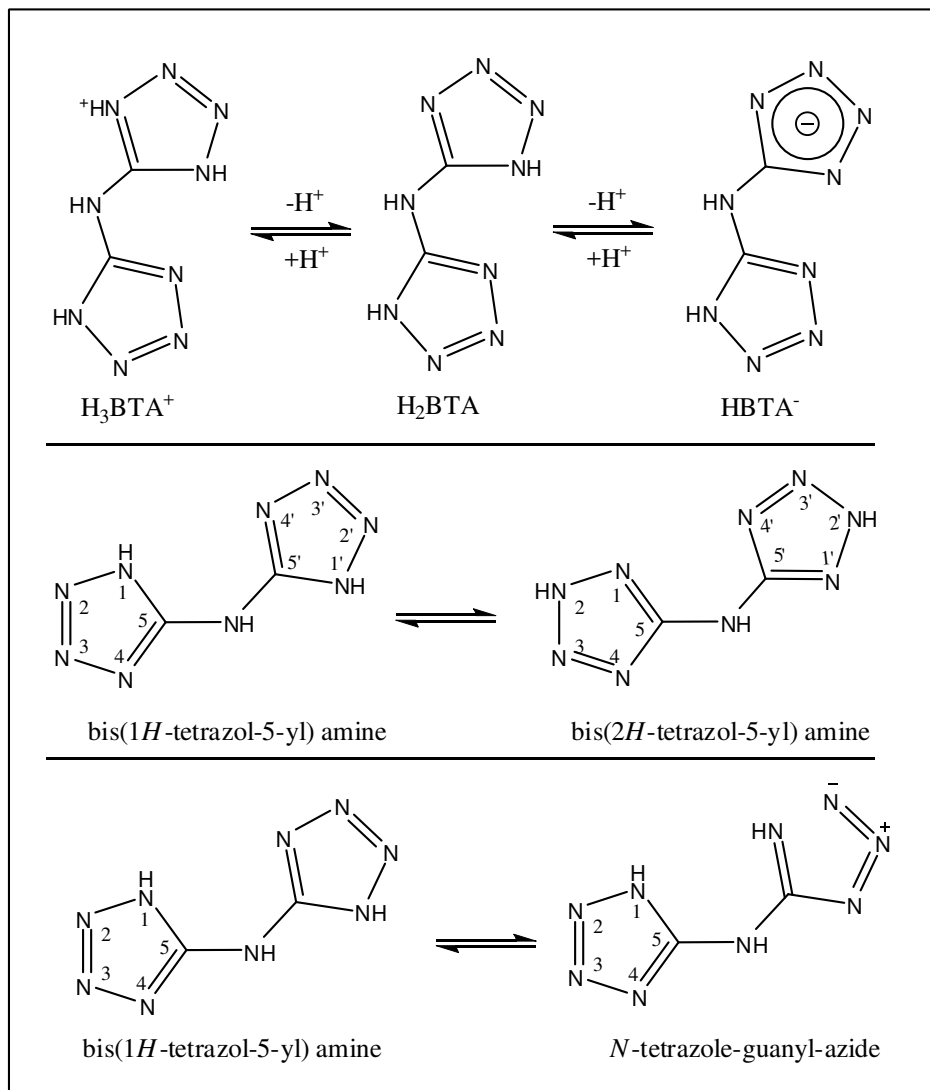
24 **Keywords:** H₂BTA; photodegradation; energetic chemical; environmental fate

25 **1. Introduction**

26 Tetrazoles are five-membered heterocyclic compounds having one carbon and four
27 nitrogen atoms arranged in planar rings. Tetrazole derivatives have wide applications in
28 major industrial areas such as medicinal chemistry, coordination chemistry, agricultural
29 technology, imaging technology, and in the automobile industry as gas generators for
30 airbags (Frija et al., 2010). Recently, they have been considered by the defense industry as
31 high nitrogen composite propellants (Klapötke et al., 2008; Yuan and Bernstein, 2016;
32 Lavoie et al., 2017). For example, in order to design a low erosivity gun propellant,
33 emphasis should be put on lowering the flame temperature and increasing nitrogen content
34 in the combustion gases. The nitrogen-rich (N% = 82.34%) bis(1*H*-tetrazol-5-yl)amine
35 (H₂BTA), an amine substituted with two tetrazole rings, fits perfectly under that role. It
36 goes without saying that new propellants designed to have lower erosivity should also
37 provide equal or better performance than the propellants they are designed to replace. The
38 positive effects of nitrogen-rich materials such as H₂BTA on the burning rate of propellants
39 and also on the gun barrel-like steel erosion have already been demonstrated by Lavoie et
40 al. (2017, 2018). These results point toward the H₂BTA becoming a very suitable candidate
41 for the future gun propellants. Although tetrazole derivatives are extensively studied,
42 H₂BTA is poorly characterized from the environmental chemistry perspective.

43 Some data on synthesis, structural attributes and isomeric distribution of the amine
44 derivatives of tetrazoles are useful to help understand the environmental chemistry and fate
45 of H₂BTA. The most common way to form tetrazoles is cycloaddition between an organic
46 nitrile (R-CN) and an organic azide (R-N₃) or azide salts. The acid-catalyzed cyclization

47 reaction of sodium dicyanamide and sodium azide in a ratio of 1:2 generates H₂BTA
48 (Klapötke et al., 2008). The hydrogen atom bound to the N-ring of tetrazole is labile and,
49 depending on the chemical environment, the molecules can be deprotonated or protonated
50 to form the corresponding anions or cations, respectively (Friedrich et al., 2005; Guo et al.,
51 2010; Li et al., 2015; Liu et al., 2017) (Fig. 1, top), or interconverted into different
52 tautomers called prototropic tautomers (Paul et al., 2009; Nelson and Baglin, 1972) (Fig.
53 1, middle). The predominant tautomeric forms of tetrazoles substituted on the carbon atom
54 (5-substituted) are mainly the 1*H*-form in the solid state and 2*H*-forms in the gas phase. In
55 nonpolar solvent, the tautomeric preference of tetrazole with an electron releasing
56 substituent (e.g. 5-aminotetrazole, 5-AT) is the same as in the gas phase, i.e., the 2*H*-form,
57 while in a solvent with high dielectric constant, like water, the 1*H*-form is dominant (Najafi
58 Chermahini et al., 2007a). The cyclic tetrazoles can also reversibly convert to an acyclic
59 azide where both forms can exist in equilibrium with each other (Fig. 1, bottom).



60

61

62 Fig. 1. Chemical structures of H₂BTA (bis(1*H*-tetrazol-5-yl)amine) and its protonated
 63 H₃BTA⁺ and deprotonated HBTA⁻ ions (top); bis(1*H*-tetrazol-5-yl)amine and its
 64 prototropic tautomer bis(2*H*-tetrazol-5-yl)amine (middle); and azido isomer (*N*-tetrazole-
 65 guanyl-azide) (bottom).

66

67 Tetrazole and its 5-substituted derivatives are weak acids with p*K*_a values in the range of
 68 1.1–6.3, which is comparable to carboxylic acids (e.g. CH₃COOH, p*K*_a = 4.75). The
 69 tetrazolic acid fragment –CN₄H has acidity similar to the carboxylic acid group, –CO₂H,

70 but is metabolically more stable at physiological pH 7.4 (Frija et al., 2010). The reported
71 p*K*_a value for tetrazole is 4.70, and two values were measured for 5-AT, 1.76 (p*K*_{a1}) and
72 5.95 (p*K*_{a2}) (Boraei, 2001). Two p*K*_a values of 1.41 (p*K*_{a1}) and 4.25 (p*K*_{a2}) have also been
73 reported for 5,5'-bitetrazole, a compound containing two tetrazole rings (Trifonov and
74 Ostrovskii, 2006). H₂BTA can be deprotonated reversibly to form mainly mono- and bi-
75 and occasionally trivalent anions. The p*K*_a values of H₂BTA have not yet been reported.
76 The anions of H₂BTA easily form salts with metals or nitrogen-rich bases such as
77 ammonium, hydrazinium and guanidium derivatives (Guo et al., 2010; Li et al., 2015; Liu
78 et al., 2008).

79 Photochemistry of mono- or bi-substituted tetrazoles (e.g. 5-methyltetrazole, 5-
80 phenyltetrazole and 2-methyl-2*H*-tetrazol-5-amine) have been widely studied, whereas
81 photodegradation of H₂BTA, the intended energetic chemical in the present study, is poorly
82 investigated. Reportedly tetrazole can undergo various photodegradation pathways
83 depending on the nature of the substitution(s) on the ring. For example, Maier and
84 coworkers (1996) reported rapid cleavage of the unsubstituted tetrazole ring during
85 photolysis, leading to the extrusion of two nitrogen atoms and the formation of several
86 photoproducts, including nitrilimine, nitrene/hydrogen cyanide (HN···HCN) complex,
87 diazomethane, carbodiimide and cyanamide. Nitrogen releases have also been reported
88 during photolysis of mono-substituted tetrazoles such as 5-phenyltetrazole in THF at
89 λ=254 nm and 5-methyltetrazole in solid argon using tunable UV laser radiation. In
90 addition, 1,2-dihydro-3,6-diphenyl-1,2,4,5-tetrazine from 5-phenyltetrazole (Scheiner and
91 Dinda Jr., 1970) and *C*-methylnitrilimine, *N*-methylcarbodiimide and methylcyanamide
92 from 5-methyltetrazole (Pagacz-Kostrzewa et al., 2014) have been identified as main

93 degradation products. During UV irradiation ($\lambda > 235$ nm) of di-substituted 2-methyl-2*H*-
94 tetrazol-5-amine in solid argon, several degradation pathways were identified including 1)
95 elimination of N₂ with the production of 1-methyl-1*H*-diazirene-3-amine, 2) cleavage of
96 the tetrazole ring leading to the production of methylazide and cyanamide, and 3)
97 tautomerization of the tetrazole to mesoionic 3-methyl-1*H*-tetrazol-3-ium-5-aminide
98 (Gomez-Zavaglia et al., 2005). The primary formation of N₂ and HN₃ has also been
99 observed during thermal degradation of 5-aminotetrazole (Brill and Ramanathan, 2000;
100 Kiselev and Gritsan, 2009; Paul et al., 2009; Piekielek and Zachariah, 2012). The photolysis
101 of H₂BTA by two photon laser excitation led to the initial cleavage of the N1-N2 bond (for
102 numbering see Fig. 1, middle) followed by the formation of N₂ (Yuan and Bernstein, 2016).
103 No other details were reported on the degradation products and other secondary
104 degradation routes.

105 The aim of this work was to provide data on the photodegradation of H₂BTA in water using
106 simulated sunlight and ultraviolet light.

107

108 **2. Materials and methods**

109 *2.1. Chemicals*

110 *N,N*-bis(1*H*-tetrazol-5-yl)amine monohydrate, H₂BTA·H₂O, was synthesized by
111 Defense Research and Development Canada (DRDC)-Valcartier according to the
112 procedures previously described by Klapötke et al. (2008). The purity of the material was
113 verified through hydrogen, carbon and nitrogen nuclear magnetic resonance spectroscopy.
114 5-Aminotetrazole monohydrate was purchased from Sigma-Aldrich (Canada). Working
115 solutions of H₂BTA (0.35 mM or 0.18 mM) and 5-AT (0.2 mM) were prepared in deionised

116 water. The pH of the H₂BTA (0.35 mM) solution was 3.65; in some assays, the pH of the
117 solution was adjusted to 5.76 and 8.05 with NaOH (0.5% w/v) prior to photolysis.

118

119 *2.2. Irradiation experiments*

120 Irradiation experiments were conducted using artificial sunlight generated from a
121 SolSim solar simulating photoreactor (Luzchem Research Inc., Canada). The total power
122 of the solar simulator output spectrum was calibrated to the best approximation of ASTM
123 Air Mass 1.5 Global Tilt Standard in the 280-800 nm regions: total irradiance of 590,000
124 mW m⁻². To accelerate the degradation process, a Rayonet photoreactor (Branford, CT)
125 equipped with 16 ultraviolet lamps (wavelength 254 nm) was also used. In a typical
126 experiment, 5 mL of aqueous solutions of H₂BTA (0.35 mM or 0.18 mM) or 5-AT (0.2
127 mM) in 20 mL quartz tubes were irradiated in either SolSim or Rayonet photoreactors at
128 room temperature. The experiments were carried out in duplicates or triplicates.

129

130 *2.3. Analytical methods*

131 H₂BTA and 5-AT were quantified by a Waters HPLC system (Milford, MA, USA)
132 equipped with a model 2996 photodiode array detector, a model 600 pump, a model 717
133 plus injector, and a temperature control module. Aliquots (10 µL) were injected on a
134 Gemini-NX-C18 column (150 x 4.6 mm, 3 µm) (Phenomenex, Torrance, CA) at 40°C. The
135 mobile phase (0.1% phosphoric acid (pH 2.2) and methanol, 90%:10%) was run
136 isocratically at 0.5 mL min⁻¹. Quantification was performed at a wavelength of 230 nm.
137 The instrumental detection limit was estimated to be 0.01 mg L⁻¹ for both chemicals.

138 The degradation products were analyzed by liquid chromatography–mass
139 spectrometry (LC-MS) using a MicrOTOF-Q mass analyzer (Bruker) attached to an HPLC
140 system (Agilent 1200 Series) equipped with a DAD detector. Aliquots (10 μ L) were
141 injected into a 3 micron-pore size Luna Phenyl-Hexyl column (2 μ m ID \times 150 mm;
142 Phenomenex, USA) at 25°C. The solvent system was composed of a mixture of MeOH and
143 0.05% v/v HCOOH in water (10%:90%) at a flow rate of 0.15 mL min⁻¹. For mass analysis,
144 negative electrospray ionization mode (ESI-) was used producing mainly deprotonated
145 molecular mass ions [M-H]⁻. The mass range was selected from 40 to 1000 *m/z*. The pH
146 measurements were carried out with a Mettler Toledo pH meter.

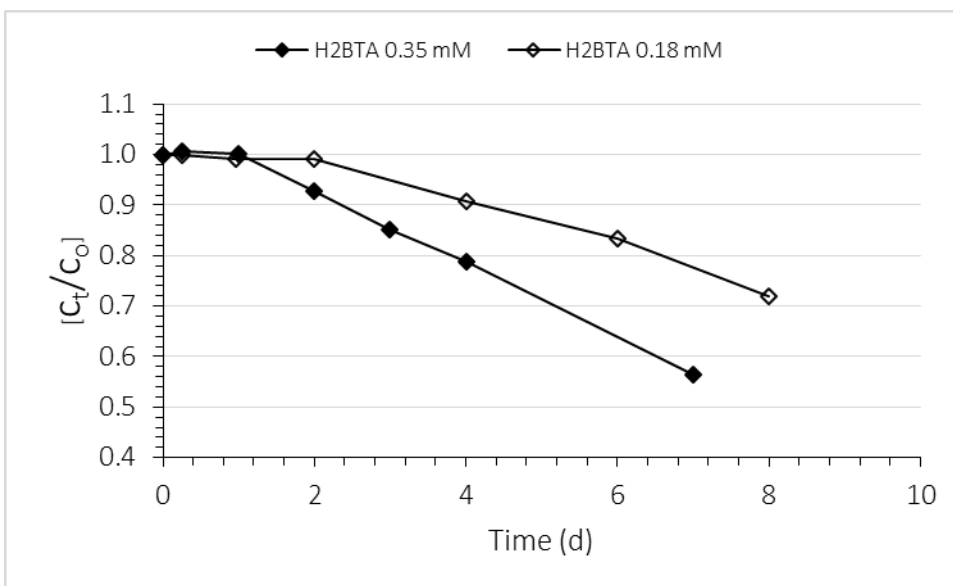
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148 **3. Results and discussion**

149 *3.1. Irradiation of H₂BTA by simulated sunlight*

150 When H₂BTA (0.35 mM, pH 3.65) was irradiated in a solar simulator (SolSim), only a
151 minute amount of the chemical disappeared after 1 day, then the H₂BTA concentration
152 decreased linearly with time (Fig. 2). A loss of 43.5% was measured after 7 day.
153 Photodegradation started immediately but very slowly for the first 2 days at the lower
154 concentration (0.18 mM, pH 3.98), reaching only 28% of degradation after 8 days of
155 irradiation (Fig. 2).

156



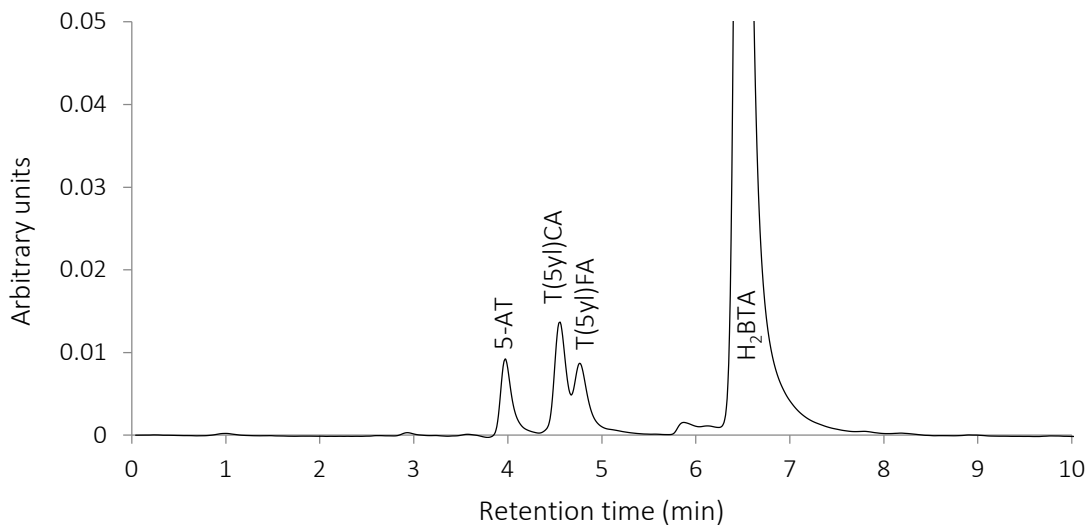
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158 Fig. 2. H₂BTA photodegradation under simulated sunlight. The disappearance of the
 159 starting material showed a RSD of less than 5%.

160

161 HPLC/UV analysis of the H₂BTA solution (0.35 mM) irradiated in the SolSim reactor
 162 showed a peak at 4.0 min that was identified as 5-AT by comparison with a commercially
 163 available reference standard (Fig. 3). Two other peaks were observed at 4.55 and 4.76 min
 164 (Fig. 3). LC-MS ESI- analysis of the compound eluted at 4.55 min showed an [M-H]⁻
 165 molecular mass ion of *m/z* 128.026 and was tentatively identified as 1*H*-tetrazol-5-
 166 ylicarbamic acid, T(5yl)CA (chemical formula: C₂H₃N₅O₂, exact molecular mass:
 167 129.029). In the same way, the compound eluted at 4.76 min showing a [M-H]⁻ molecular
 168 mass ion *m/z* 112.030 was tentatively identified as *N*-(1*H*-tetrazol-5-yl)formamide,
 169 T(5yl)FA (chemical formula: C₂H₃N₅O, exact molecular mass: 113.034).

170



171

172 Fig. 3. HPLC/UV chromatogram of H₂BTA (0.35 mM, pH 3.65) photolyzed in a solar
 173 simulator for 7 days.

174

175 While trace amounts of T(5yl)FA and T(5yl)CA were detected after 1 day, formation of 5-
 176 AT did not begin until day 2 (Fig. S1), suggesting that T(5yl)FA (*N*-(1*H*-tetrazol-5-
 177 yl)formamide) and T(5yl)CA (1*H*-tetrazol-5-ylcarbamic acid) could be the precursors of
 178 5-AT. The highest yields of 5-AT, 0.43 and 0.39 moles for each mole of H₂BTA degraded
 179 (using 0.35 mM and 0.18 mM, respectively), were measured after 2 days of irradiation and
 180 then decreased to 0.26 moles in both cases after 7 and 8 days of photolysis in SolSim,
 181 respectively. Due to the demonstration in a separate experiment of the very slow
 182 degradation of 5-AT (0.2 mM) by artificial sunlight (only 2.5% degraded after 8 days), we
 183 suspect that the transformation of one or more of the presumed precursors of 5-AT occurred
 184 at a slower rate than their formation.

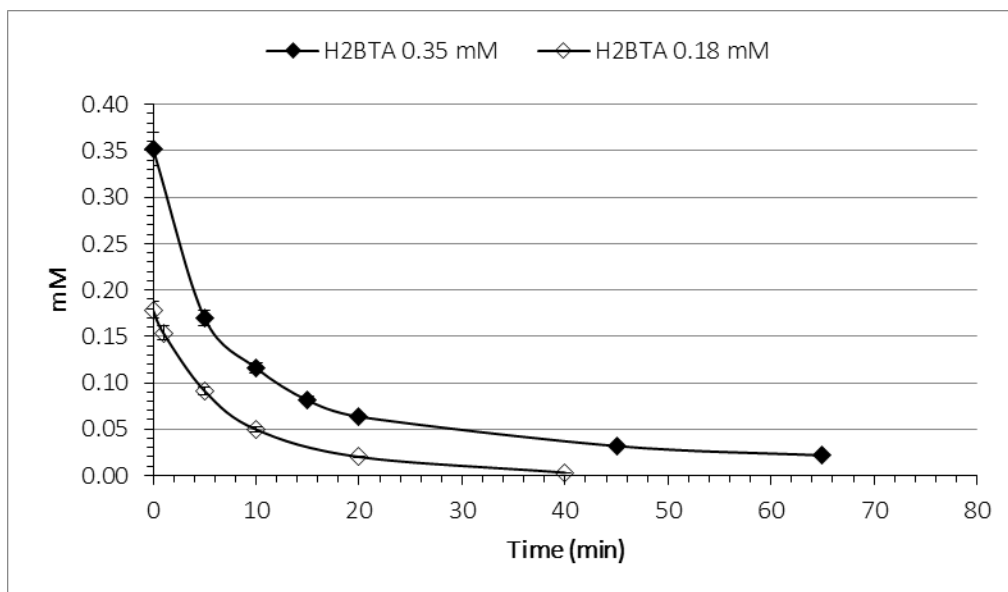
185 The initial pH (3.65) of the 0.35 mM H₂BTA solution decreased to 3.50 after 4 days
186 of irradiation in SolSim, suggesting slow deprotonation of H₂BTA to form HBTA⁻ anions.
187 Since pK_a values of H₂BTA in water are unknown, it can only be estimated that at pH 3.65,
188 H₂BTA is poorly dissociated in water, but its dissociation might increase by the influence
189 of light energy. Interestingly, in the presence of a greater amount of HBTA⁻ anions, e.g. at
190 pH 5.76 adjusted with a small amount of NaOH (0.5%), photodegradation started
191 immediately without a lag phase (Fig. S2), suggesting that HBTA⁻ anions could be involved
192 in the initial photodegradation of H₂BTA. However, at pH 8.0, H₂BTA did not degrade
193 after 3 days of irradiation with simulated sunlight. We hypothesize that, at this pH,
194 dissociation of H₂BTA is more pronounced, leading to the formation of the dianion BTA²⁻
195 with a higher aromaticity, and thus more stability, than the HBTA⁻ anion.

196

197 *3.2. Irradiation of H₂BTA by UV at λ=254 nm*

198 H₂BTA in water (0.35 mM, pH 3.65 and 0.18 mM, pH 3.98) irradiated in a Rayonet
199 photoreactor at λ=254 nm disappeared rapidly, resulting in a loss of 94% after 65 min and
200 of 98% after 40 min, respectively (Fig. 4). The plot of {1/[c]_t} vs time for the 0.35 mM
201 experiment gave a straight line with slope k = 0.666 mM⁻¹ min⁻¹ (r² = 0.999), indicating that
202 photolysis of H₂BTA proceeded with second-order kinetics (Fig. S3). A half-life of 4.29
203 min was calculated. On the other hand, plotting {ln[c]_t} vs time for the 0.18 mM
204 experiment gave a straight line with slope k = -0.1 min⁻¹ (r² = 0.995) indicating that
205 photolysis of H₂BTA proceeded via a first-order kinetics (Fig. S4). A half-life of 6.92 min
206 was calculated. Changing the order of a chemical reaction by changing the initial

207 concentration of the reactant is an indication that the reaction proceeds via mixed-order
208 kinetics, a characteristic of a complex (multi-step) reaction.
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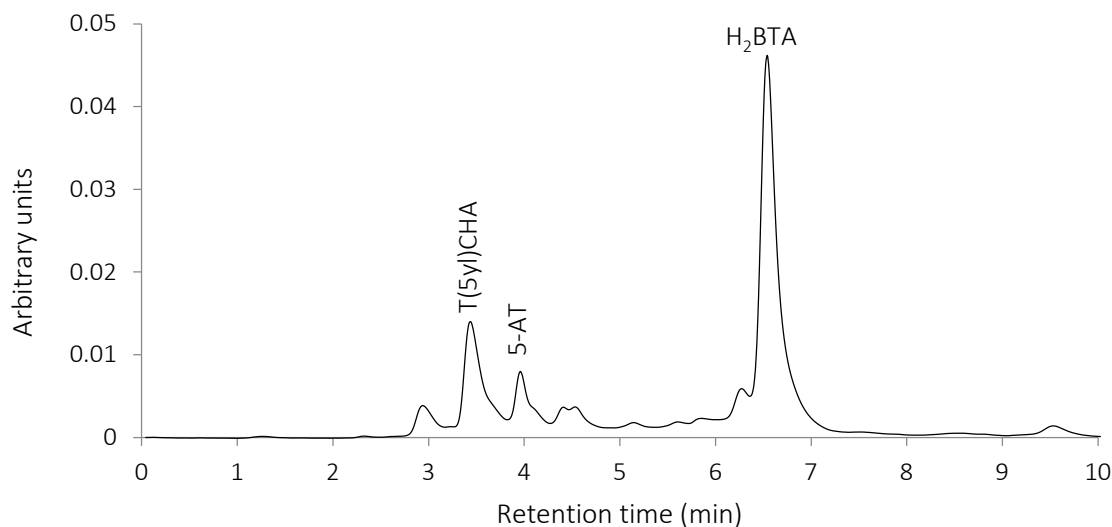
211 Fig. 4. Degradation of H₂BTA in water (0.35 mM, pH 3.65 and 0.18 mM, pH 3.98) during
212 irradiation at $\lambda = 254$ nm. The disappearance of the starting material showed a RSD of less
213 than 5%.

214

215 During degradation of H₂BTA (0.35 mM) with UV light, production of 5-AT was
216 observed, but T(5yl)CA was not detected. T(5yl)FA appeared only in traces after 40 min
217 of irradiation. Instead, HPLC/UV showed a new peak at 3.45 min (Fig. 5) that was
218 tentatively identified by LC-MS as *N*-(1*H*-tetrazol-5-yl)carbamohydrazonic acid,
219 T(5yl)CHA (chemical formula: C₂H₅N₇O, exact molecular mass: 143.055), a water adduct
220 of nitrile imine derivative produced by the photochemical extrusion of N₂ from H₂BTA.
221 The photoinduced rupture of the tetrazole ring ($\lambda < 290$ nm) and water addition to the 1,3-
222 dipolar nitrile imine generated was reported by Herner and Lin (2016). Quenching of the

223 nitrile imine intermediate by H₂O produced during photolysis of a peptide tetrazole at
224 $\lambda=302$ nm was also described by Song et al. (2008).

225



226

227 Fig. 5. HPLC/UV chromatogram of H₂BTA (0.35 mM, pH 3.65) photolyzed at $\lambda = 254$ nm
228 for 20 min.

229

230 Formation of T(5yl)CHA peaked between 10 min and 20 min then slowly decreased;
231 however, 5-AT remained almost constant (Fig. S5). The yields of 5-AT were only 0.06 and
232 0.09 moles for each mole of degraded H₂BTA, for initial concentrations of 0.35 and 0.18
233 mM, respectively, after 40 min of irradiation. When 5-AT (0.2 mM) was irradiated in a
234 separate experiment, 91% of the compound was degraded after 40 min, slightly less than
235 the degradation of H₂BTA (98% of 0.18 mM) under the same conditions, suggesting that
236 5-AT was produced and transformed with almost similar rates, which explains the low 5-
237 AT yields observed during photolysis of H₂BTA at $\lambda = 254$ nm.

238

239 3.3. H₂BTA photodegradation pathways

240 The photodegradation of H₂BTA at pH 3.65 by simulated sunlight began with a relatively
241 long lag phase (1 day). In addition, the pH decreased slightly, which can be explained by
242 the deprotonation of H₂BTA, forming HBTA⁻ anions. An increase in the initial degradation
243 rate was observed at pH 5.76 when irradiated with simulated sunlight, suggesting that
244 HBTA⁻ anions are involved in the photodegradation of H₂BTA. Detection of the T(5yl)FA
245 and T(5yl)CA intermediates and 5-AT indicates that only one tetrazole ring was cleaved
246 during photolysis of H₂BTA by artificial sunlight (Scheme 1; pathway *a*). We hypothesize
247 that deprotonation of the ring-NH group might have modified the ring aromaticity thus
248 facilitating isomerization, i.e., azido-tetrazole rearrangement, of the second tetrazole ring
249 leading to the loss of HN₃. The loss of HN₃ from the anion forms of an 1*H*-tetrazol-5-yl
250 derivative, N₃,N₆-bis(1*H*-tetrazol-5-yl)-1,2,4,5-tetrazine-3,6-diamine (Fig. S6), during
251 ESI- mass spectrometry measurement has been reported previously (Parvari et al., 2018).
252 The authors found that the activation energy for HN₃ loss is considerably lower than those
253 of the proton-migration-dependant fragmentation leading to loss of N₂ (Parvari et al. 2018).
254 When one of the two tetrazole rings of H₂BTA is destroyed, the 5-amino group can stabilise
255 the second tetrazole ring (Cubero et al., 1998), resulting in greater resistance of 5-AT to
256 photolysis with simulated sunlight.

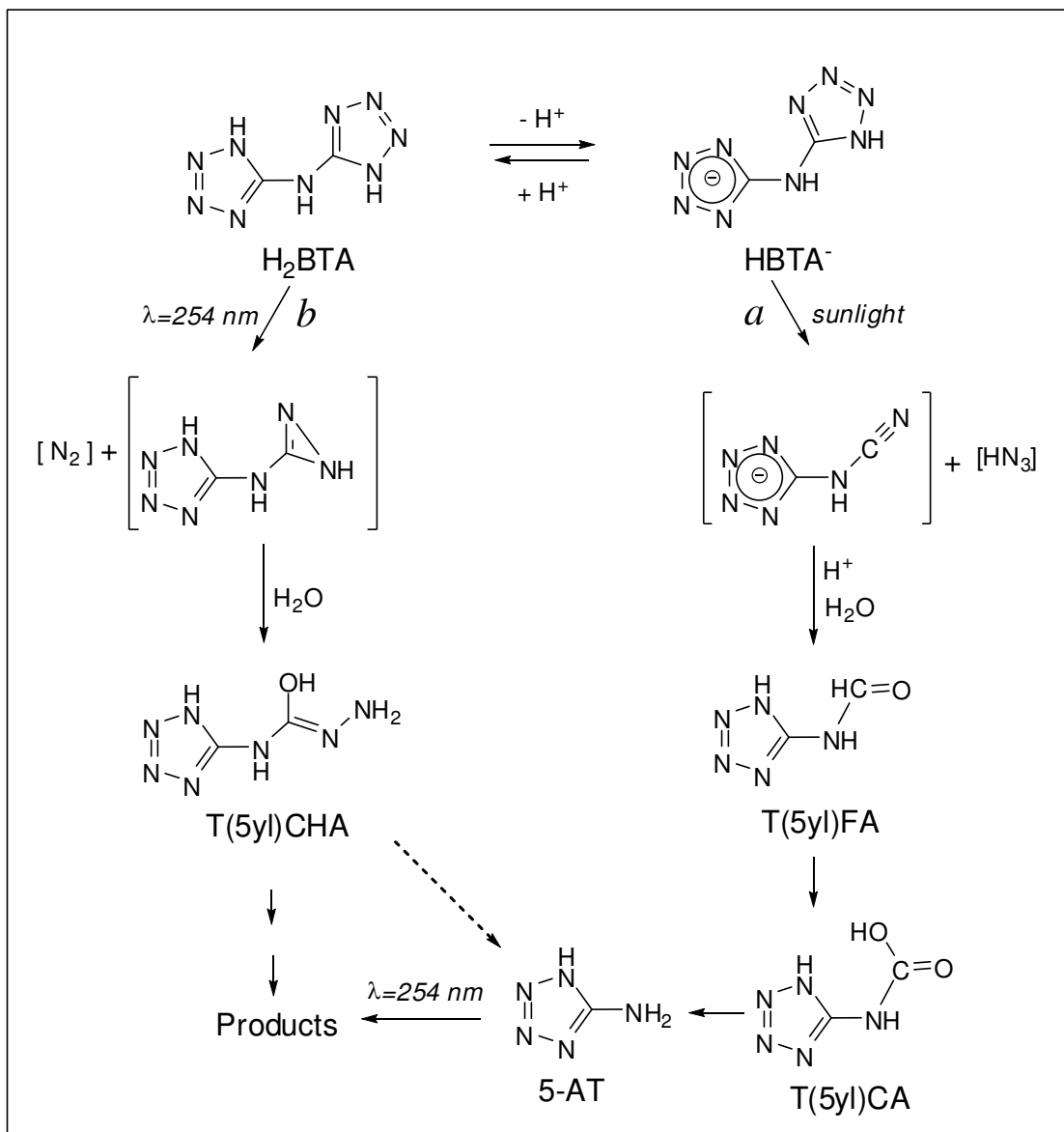
257 Compound T(5yl)CHA was only detected in samples irradiated with UV light ($\lambda =$
258 254 nm), suggesting that a second pathway also occurred under this condition. H₂BTA can
259 also be degraded by extruding N₂ from one of the two tetrazole rings producing the
260 presumed nitrile imine-H₂O adduct T(5yl)CHA (Scheme 1; pathway *b*). In situ formation

261 of nitrile imine upon UV irradiation of tetrazoles derivatives and its reactivity with water
262 has been reported (Song et al. 2008; Herner and Lin, 2017), but never observed during
263 photolysis of the energetic chemical H₂BTA in water. T(5yl)CHA was unstable and
264 degraded like 5-AT to unidentified products (Scheme 1; pathway *b*).

265

266

267



268

269 Scheme 1. Postulated main degradation pathways during photolysis of H₂BTA in water; a)
 270 with simulated sunlight and b) with UV (λ=254 nm).

271

272 The possibility of transforming T(5yl)CHA into intermediate 5-AT cannot be ruled out;
 273 however, LC-MS detection of trace amounts of T(5yl)FA in samples irradiated for 40 min

274 and 65 min suggests that both pathways *a* and *b* occurred simultaneously (Scheme 1)
275 during UV irradiation.

276 **Conclusion**

277 The nitrogen-rich compound H₂BTA was partially degraded in water under solar
278 simulating conditions and completely with UV light. The kinetic study and products
279 analysis revealed that photo-transformation of H₂BTA with simulated sunlight was
280 governed by a major pathway; however, a second degradation path appeared
281 simultaneously during UV irradiation ($\lambda = 254$ nm). We proposed a photodegradation
282 pathway triggered by two independent initial routes: a) one initiated by the loss of HN₃ and
283 b) another initiated by the loss of N₂ to eventually give 5-AT. 5-AT remained relatively
284 stable under simulated sunlight, but was degraded to unidentified products by UV light.
285 Apart from the present study showing the potential photoremoval and thus natural
286 attenuation of H₂BTA in contaminated water bodies, no other information is available on
287 the fate and toxicity of H₂BTA in the environment. It is thus worth investigating whether
288 5-AT, a key degradation product of H₂BTA, is more or less toxic than the parent compound.
289 As this chemical has a high probability of being used in future insensitive and high energy
290 propellant formulations, the study of its environmental fate will contribute to a risk
291 assessment prior to deployment in ranges and training areas.

292

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296 valuable review of the manuscript.

297

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Supporting Information

Photolysis of bis(1*H*-tetrazol-5-yl)amine (H₂BTA), a high nitrogen content tetrazole-based energetic compound in water

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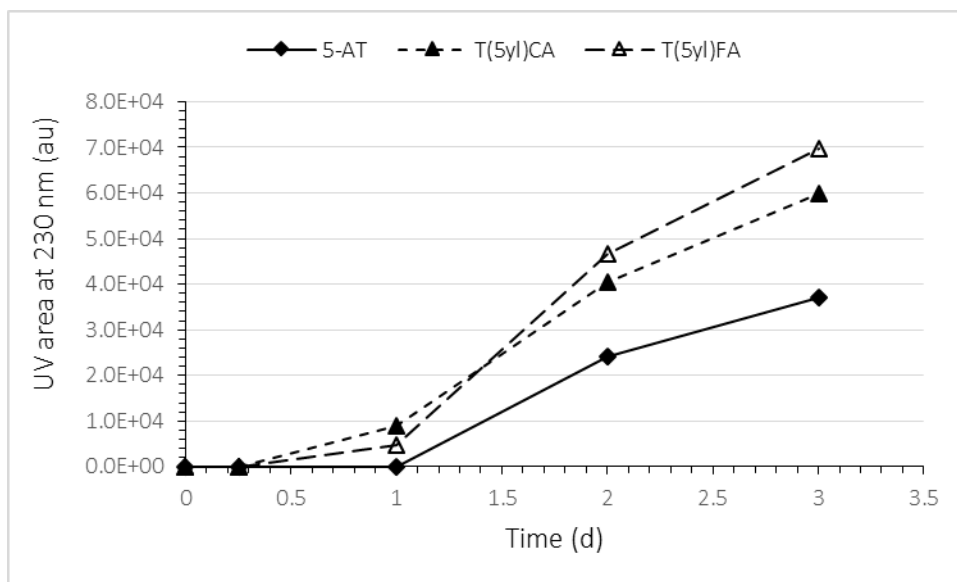


Figure S1. Degradation products of H₂BTA (0.35 mM, pH 3.65) irradiated by simulated sunlight.

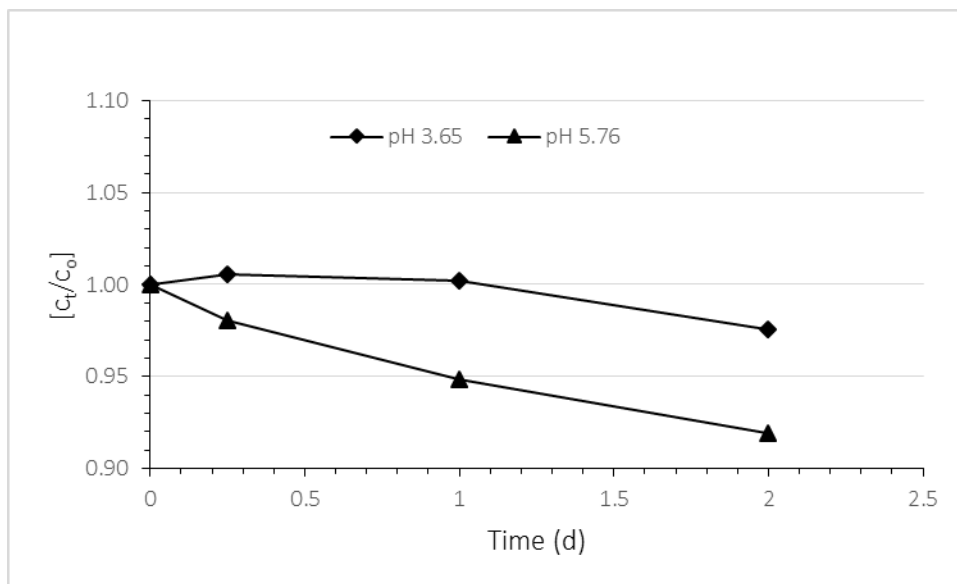


Figure S2. Initial degradation of H₂BTA irradiated in a SolSim reactor at pH 3.65 and pH 5.76.

Errors bars of duplicate experiments (< 5%) are not presented.

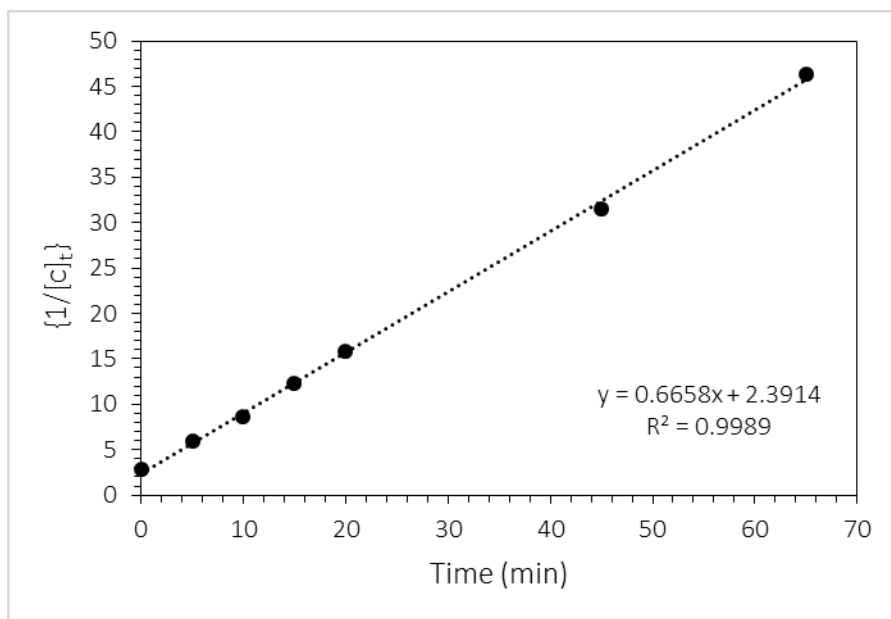


Figure S3. Plot of $\{1/[c]_t\}$ vs time of H₂BTA degradation (0.35 mM in water, pH 3.65) during UV ($\lambda=254$ nm) irradiation. RSD of duplicate experiments were less than 5%.

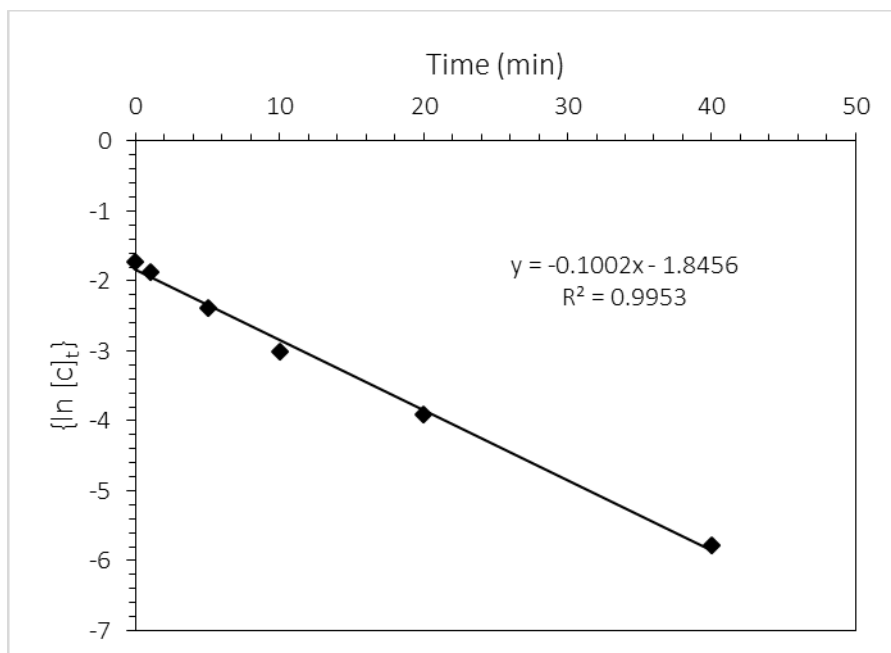


Figure S4. Plot of $\{\ln [c]_t\}$ vs time of H₂BTA degradation (0.18 mM in water) during UV ($\lambda=254$ nm) irradiation. RSD of duplicate experiments were less than 5%.

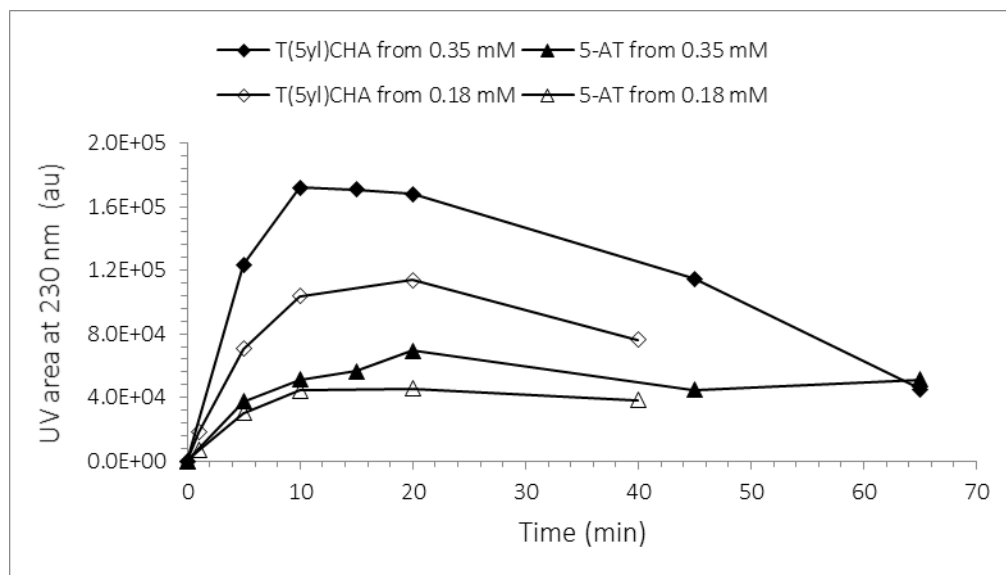


Figure S5. Degradation products of H₂BTA (0.35 mM and 0.18 mM, pH 3.65) photolyzed at $\lambda = 254$ nm.

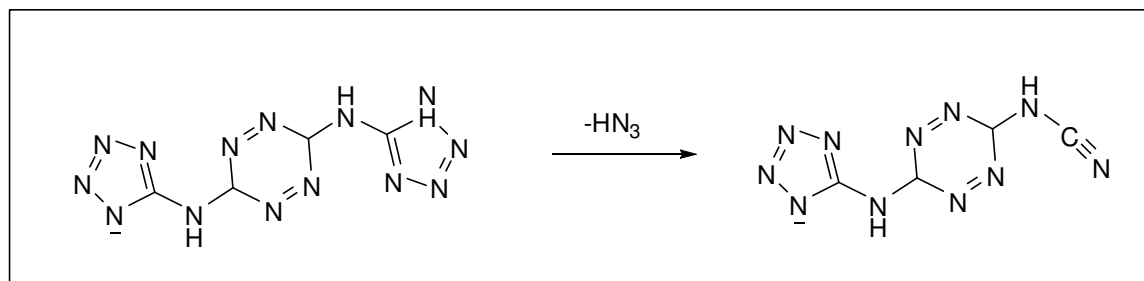


Figure S6. The loss of HN₃ from the anion forms of N₃,N₆-bis(1*H*-tetrazol-5-yl)-1,2,4,5-tetrazine-3,6-diamine proposed by Parvari et al. (2018).